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# Satellite observation of hourly dynamic characteristics of algae with Geostationary Ocean Color Imager (GOCI) data in Lake Taihu



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# ABSTRACT

Phytoplankton bloom in a shallow inland eutrophic lake (Taihu Lake) is characterized by significant spatial and temporal variation and a high concentration of chlorophyll-a ( $C_{chl-a}$ ). The observation of the rapidly changing dynamic characteristics of algae is limited by the insufficient temporal resolution of satellite data. The Geostationary Ocean Color Imager (GOCI), launched by Korea, can provide high temporal resolution satellite data to observe the hourly dynamics of algae. In this study, a simple regional NIR-red two-band empirical algorithm of  $C_{chl-a}$  for GOCI is proposed for Taihu Lake. Study results show that the GOCI-derived  $C_{chl-a}$  matches the in situ measured values well. Based on this validated algorithm of  $C_{chl-a}$ , we obtained the hourly maps of  $C_{chl-a}$  from GOCI level-1b data during the period August 6 to August 9, 2013. The spatial variation of GOCI-derived  $C_{chl-a}$  also matches synchronous in situ measured values well, and the temporal variation of GOCI-derived  $C_{chl-a}$  also matches synchronous in situ measured values well, and the temporal variation of GOCI-derived  $C_{chl-a}$  in Taihu Lake. The vertical current plays an important role in the hourly scale of spatial and temporal variations in phytoplankton to the distribution of phytoplankton over the long term, but spatially and temporal variations in the short term.

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# 1. Introduction

Algae are vital to marine and freshwater ecosystems because of their key role in keeping the biosphere balanced despite natural forces and food-chain relationships (Cloern, 2001; Rabalais, 2004). Algal blooms, however, which are formed from the excessive growth of algae in freshwater and coastal marine ecosystems, are caused by increased nutrients (i.e., nitrogen and phosphorus) and have become a global problem. Algal blooms (such as *Microcystis aeruginosa, Scenedesmus obliquus* and diatom) occur worldwide. Taihu, Dianchi and Chaohu Lakes in China (Huang, Li, Yang, Sun, et al., 2014; Huang, Wang, et al., 2014; Paerl et al., 2011), Lake Erie in the United States (Michalak et al., 2013; Stumpf, Wynne, Baker, & Fahnenstiel, 2012), Lakes Wood and Winnipeg in Canada (Binding, Greenberg, & Bukata, 2011; Schindler, Hecky, & McCullough, 2012), and Lake Nieuwe Meer in the Netherlands (Jöhnk

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et al., 2008), have received much attention for their harmful toxins. The blooms are generally perceived to result from a rapid increase in biomass, because the algae growth period is much longer than the loss period. Previous studies have proposed that the intensity and frequency of blooms are empirically connected to nutrient status (Abell, Ozkundakci, & Hamilton, 2010; Paerl et al., 2011; Schindler et al., 2008; Xu, Paerl, Qin, Zhu, & Gao, 2010). Temperature, light and water dynamic characteristics are also important influencing factors (Carstensen, Henriksen, & Heiskanen, 2007; Moore et al., 2008; Paerl & Huisman, 2009; Wong, Lee, & Hodgkiss, 2007; Zhang, Duan, Shi, Yu, & Kong, 2012). The formation process of algal bloom is clearly complex.

Indeed, regulating nutrient input is the only realistic method of controlling algal bloom intensity and frequency (Conley et al., 2009; Paerl et al., 2011). As an extremely eutrophic lake, the nutrient input into the Taihu Lake watershed was effectively controlled, but the dynamic release of nutrients from sediment was not. (Dzialowski, Wang, Lim, Beury, & Huggins, 2008; Qin, Xu, Wu, Luo, & Zhang, 2007; Qin et al., 2006). Consequently, eutrophication could not be controlled in a short time period. However, the high rates of algal blooms create an urgent need for effective measures to reduce loss caused by blooms.



Fig. 1. Study area. The bottom right corner is the location of Taihu Lake in Jiangsu province. The black star is the position of buoy. The background of Taihu Lake is thickness of sediment. Taihu Lake can be separated into seven segments, including: Meiliang Bay (MB), Central Lake (CL), Gonghu Lake (GL), Western Lake (WL), Southwestern Lake (SWL), Eastern Lake (EL) and Eastern Bay (EB), Zhushan Bay (ZSB). A is a photo of the buoy. The main portion of the Eastern Bay is characterized by emergent aquatic plants, such as those shown in chart B.

Monitoring algal bloom and chlorophyll-a to evaluate and predict algal bloom is thus a critical prerequisite to finding effective methods to control the bloom. Satellite remote sensing provides rapid, synoptic, and repeated information on water quality (Duan et al., 2009; Hunter, Tyler, Willby, & Gilvear, 2008a, 2008b; Liu, Wang, & Shi, 2009; Odermatt et al., 2012). Many retrieval algorithms, such as the float algal index (Gower, King, Borstad, & Brown, 2005; Hu, 2009; Hu et al., 2005; Wynne, Stumpf, Tomlinson, & Dyble, 2010) and new chlorophyll-a retrieval models (Huang, Li, Yang, Li, et al., 2014; Shanmugam, 2011), have been developed to study the algal bloom, biological, and ecological processes and phenomena in Taihu Lake (Duan et al., 2009; Hu et al., 2010; Huang, Li, Yang, Li, et al., 2014; Le et al., 2009). Long-term records of algal blooms in the lake based on satellite data have been established to reveal large temporal and spatial variation of algal bloom (Duan et al., 2009; Hu et al., 2010; Wang, Shi, & Tang, 2011). However, there has been little published work on the dynamic characteristics of algae observation based on satellite data. This is due to the challenge of obtaining high temporal resolution, accurate atmospheric correction algorithms and a chlorophyll-a retrieval model for the satellite data. Hydrodynamic force has a significant effect on the formation of algal blooms and on their distribution (Moreno-Ostos, Cruz-Pizarro, Basanta, & George, 2009; Wu & Kong, 2009; Wu et al., 2013). This is particularly true for Taihu Lake because it is a shallow inland lake with a high dynamic ratio ([square root of the area]/depth: 25.4) (Huang, Li, Yang, Sun, et al., 2014), where algal (mainly is *M. aentginosa*) blooms occur every year (Guo, 2007; Huang, Li, Sun, & Le, 2011). The observation of the dynamic characteristics of algae in Taihu Lake using high temporal resolution satellite data is thus necessary and important. The Geostationary Ocean Color Imager (GOCI), launched by Korea, was decided on as a good choice for this observation (Choi et al., 2012; He et al., 2013; Ruddick et al., 2012; Ryu, Choi, Eom, & Ahn, 2011).

In the study, we establish a simple NIR-red band ratio algorithm of chlorophyll-a concentration ( $C_{chl-a}$ ) for GOCI data using in situ measurements of remote sensing reflectance and  $C_{chl-a}$ , and reveal the dynamic characteristics of  $C_{chl-a}$  and phytoplankton bloom by retrieval results of  $C_{chl-a}$  from GOCI. This provides reliable satellite observation data to better understand the dynamics of phytoplankton bloom and the factors controlling bloom in Taihu Lake.

# 2. Data and methods

#### 2.1. Study area

Taihu Lake (30°90′–31°54′N and 119°55.3′–120°59.6′E) is located on the Yangtze River delta, and is a very important drinking water source for the cities of Suzhou, Wuxi and Shanghai. It is the thirdlargest freshwater lake in China with an area of 2428 km<sup>2</sup> (water surface area is 2338 km<sup>2</sup>, island area is 90 km<sup>2</sup>), and a mean depth of 1.9 m (Fig. 1). It is influenced by the East Asian monsoon climate. The lake can be separated into seven major segments, including: Meiliang Bay



Fig. 2. Calibration and validation of C<sub>chl-a</sub>NIR-red two-band algorithm.



Fig. 3. Atmospheric correction results by 6S model with real-time meteorological parameters. M is measured data, AC is atmospheric correction result, and numbers with µg/L units are  $C_{\rm chl-a}$ .



**Fig. 4.** A is validation of *C*<sub>chl-a</sub>NIR-red two-band algorithm by match-up points, B is the synchronous GOCI data on 05/12, 05/13 and 08/01 with the match-up points. The clay in B 2013/08/01/00 is cloud. The RGB image combined from three bands of 745 nm, 555 nm and 412 nm.

(MB), Central Lake (CL), Western Lake (WL), Southwestern Lake (SWL), Gonghu Lake (GL), Eastern Lake (EL) and Eastern Bay (EB) (Fig. 1). The lake became increasingly eutrophic during the 1980s, and its current key issues are hyper-eutrophication and algal blooms (Guo, 2007; Wang and Shi, 2008). The area of blooms has increased significantly, the blooms' duration has lengthened, and their initial blooming date has become earlier (Duan et al., 2009).

# 2.2. Data sources

#### 2.2.1. Satellite image and data process

GOCI satellite image data with 500 m spatial resolution and 1-hour temporal resolution were used to get the hourly map of  $C_{chl-a}$ . Match up data, GOCI level-1b, in 05/12/2013, 05/13/2013, 08/01/2013 and from 08/05/2013 to 08/09/2013 (full clear-sky scenes) were obtained from the Korea Ocean Satellite Center (http://kosc.kordi.re.kr/). The technical software for GOCI (GDPS) is specialized for the ocean, and land information was masked during data processing. Thus, GOCI level-1b data were converted to calibrated radiance data using the self-compiled ENVI–IDL (Interface description language) software package, in which all calibration parameters are from a hierarchical data of the GOCI image. A geometric correction is also included in the software package. Atmospheric influence was corrected (atmospheric correction) based on the radiative transfer calculations from the Second Simulation of the Satellite Signal in the Solar Spectrum (6S model) (Vermote et al., 1997). The continental aerosol type and atmospheric profiles (middle latitude summer type) embedded in the 6S model was used. The MODIS product of aerosol optical thickness surrounding the Taihu Lake was set as input parameters to the 6S model (aerosol concentration). Firstly, the spatial statistic (three pixels around Taihu Lake) mean value of MODIS product of aerosol was calculated and then this mean value (represent the aerosol concentration of whole lake, this process is validated for the full clear-sky scenes) was used to input into 6S model (aerosol concentration in item). Some small improvements to the 6S model were carried out as well, including the addition of real-time atmospheric humidity and pressure. The output parameters from 6S were put into the following equations and calculated the corrected reflectance

$$y = X_a * L - X_b$$
  
$$R_{rs} = y/(1 + X_c * y)$$

where  $X_a$ ,  $X_b$  and  $X_c$  are the output parameters from 6S model,  $R_{rs}$  is the atmospheric corrected reflectance, and y is the process variable.

#### 2.2.2. In situ measurement

The in situ measurements of remote sensing reflectance  $[R_{rs}(\lambda), Sr^{-1}]$  and chlorophyll-a concentration ( $C_{chl-a}, \mu g/L$ ) were conducted from 2004 to 2013. Field measurements, including 1228 samples from 38 cruises, were carried out. The in situ measurement during the period 08/05 to 08/09 in 2013 also contains information on algal species, which were identified with a microscope by manual visual judgment.

 $R_{rs}(\lambda)$  measurements were obtained using an analytical spectral device called FieldSpec spectroradiometer (350–1050 nm range, with a

**Fig. 5.** Spatial variations of  $C_{chl-a}$  mapped by GOCI data during daytime hours from August 6 to August 9. The odd rows are RGB images from three bands of 745 nm, 555 nm and 412 nm. The even rows are GOCI-derived  $C_{chl-a}$ . The range of color bar was set to 0–200 µg/L, the GOCI-derived  $C_{chl-a}$  with crimson can be considered algal bloom. Eastern Lake and Eastern Bay were excluded in our analysis so as not to affect our analysis for other segments. For the completeness and aesthetics of figure, areas of Eastern Lake and Eastern Bay were not masked in this study.



sampling interval of ~1 nm). The spectroradiometer was calibrated by equipment manufacturers every year.  $R_{rs}(\lambda)$  was obtained from the measured reflectance radiance of a standard reflectance panel  $[L_p(\lambda)]$ , the upwelling radiance of water  $[L_{tw}(\lambda)]$  and the downwelling radiance of sky  $[L_s(\lambda)]$ . The influence factors (such as white hat bubble and solar flare) to measurement of radiance were fully considered and avoided during the period of in situ measurements. Ten spectra were observed for each target and abnormal spectra were removed. The mean value of the rest of the spectra was used to derive  $R_{rs}(\lambda)$  (Huang et al., 2011).

Water samples were filtered with GF/C filters (Whatman). The filtered GF/C filters were used to extract chlorophyll-*a* with ethanol (90%) at 80 °C for 6 h in darkness and then analyzed by Shimadzu UV-3600 (Huang, Zou, et al., 2014c). Absorbance difference at 750 and 665 nm, which are before and after removal of phaeopigments by adding hydrochloric acid with 1 mol/L, was used to calculate  $C_{chl-a}$  (Chen, Chen, & Hu, 2006).

# 2.2.3. Real-time observation data by buoy

The buoy located in Meiliang Bay (marked by a black star in Fig. 1) measured real-time and consecutive data from 08/06/2013 to 08/09/2013 with fifteen-minute intervals. The data include meteorological data (wind speed, wind direction) (Airmar-PB200, USA) and water-current data for different floors (YSI-SonTEK, ADP-1000-1 MHz, USA) and  $C_{chl-a}$ . The  $C_{chl-a}$  was measured by fluorometer with a measuring range of 0–100 relative fluorescence units (YSI6600V2-4, USA) at the depths of 0.4 m and 0.8 m.

# 2.2.4. Application of field measured data

The field measured data was used to calibrate and validate the C<sub>chl-a</sub> NIR-red two-band algorithm. In order to get the relatively more robust model, most of measurement data, which included 1126 samples (24 samples with phytoplankton bloom were removed), was used to calibrate the retrieval algorithm of  $C_{chl-a}$ . The data (78 samples), collected in summer (2013), were used to validate the retrieval algorithm because the algorithm will be applied to aestival GOCI satellite data. Among of them, seven days of match-up points were found in the data, namely, 05/12, 05/13, 08/01, 08/05, 08/06, 08/07, 08/08 and 08/ 09 in 2013. Three days of match-up samples (05/12, 05/13 and 08/01) were used to validate the  $C_{chl-a}$  NIR-red two-band algorithm and atmospheric correction method for the GOCI data. The match-up points during the period of 08/05 to 08/09 were used to confirm the validity of GOCI-derived C<sub>chl-a</sub> from 08/06/2013 to 08/09/2013 and ensure the correctness of the interpretation of retrieval result. The buoy-observed C<sub>chl</sub>a data during the period of 08/05 to 08/09 was used to validate the capture capability of GOCI for the dynamics of phytoplankton.

#### 2.3. Accuracy assessments

The difference between measured and retrieved value was also normalized to the measured value, called the root mean square error percentage (RMSP). This RMSP was used to assess the accuracy of the NIR-red two-band algorithm. The calculation equation is

$$RMSP = \sqrt{\frac{\sum_{i=1}^{n} \left[ (C_{chl-ar} - C_{chl-am}) / C_{chl-am} \right]^2}{n}}$$

where  $C_{chl-ar}$  and  $C_{chl-am}$  are the retrieved and measured  $C_{chl-ar}$ 

# 3. Retrieval model and validation

3.1. Calibration of  $C_{chl-a}$  NIR-red two-band algorithm for simulated GOCI bands

It has been demonstrated that an empirical  $C_{chl-a}$  NIR-red algorithm can be used for estimating  $C_{chl-a}$  in highly turbid productive waters with

satisfactory performance (Huang, Zou, et al., 2014; Gilerson et al., 2010; Le et al., 2013; Zhang et al., 2011). Targeted at the NIR-red bands of GOCI (745 nm, 680 nm and 660 nm), a simple NIR-red two-band algorithm was established. The in situ  $C_{chl-a}$  and  $R_{rs}(\lambda)$ , measured during 37 cruises (1126 points), shows a good relationship between  $C_{chl-a}$  and  $R_{rs}(745)/R_{rs}(680)$  at the GOCI bands, which is the average value of  $R_{rs}(\lambda)$  according to the width of the GOCI bands (745  $\pm$  10 nm and 680  $\pm$  5 nm) (Huang, Zou, et al., 2014). A total of 1126 samples were used to calibrate the  $C_{chl-a}$  NIR-red two-band algorithm, and 78 samples were used to validate the  $C_{chl-a}$  NIR-red two-band algorithm. The calibration algorithm of  $C_{chl-a}$  is

$$C_{chl-a} = 10^{F(ratio)}, (R^2 = 0.71, N = 1126, p < 0.001)$$
  
 $F(ratio) = 1.8875 + 0.8296 * (1 - 0.2241^{\log(Rrs(745)/Rrs(680), 10)})$ 

where,  $R^2$  is the coefficient of determination, N is the sampling number. Almost all the samples (>98%) were in the 95% prediction band (Fig. 2A). The RMSP of validation result between measured and retrieved  $C_{chl-a}$  is 36.44% (Fig. 2B). The test result indicates that the  $C_{chl-a}$  NIR-red two-band algorithm can estimate  $C_{chl-a}$  with acceptable accuracy (Shafique, Autrey, & Fulk, 2001).

# 3.2. Validation of C<sub>chl-a</sub> NIR-red two-band algorithm for GOCI image

# 3.2.1. Validation of atmospheric correction

To examine the applicability of the C<sub>chl-a</sub> NIR-red two-band algorithm to GOCI satellite images, the match-up data (05/12, 05/13, 08/01) in 2013 are used to validate the NIR-red two-band algorithm and atmospheric correction method for the GOCI data. The atmospheric correction results for the match-up points of GOCI show that the performance of the 6S model with real-time meteorological parameters is encouraging, although there is still some uncertainty (Fig. 3). The maximum and minimum relative errors between the measured and atmospheric correction results are -40.65% for  $R_{\rm rs}(745)$  with 51.15 µg/L  $C_{\rm chl-a}$  and 3.76% for  $R_{\rm rs}(680)$ with 9.77 µg/L C<sub>chl-a</sub> (blue line in Fig. 3A and green line in Fig. 3B, respectively). The RMSP of measured and atmospheric correction results  $[R_{rs}(745)]$  and  $R_{rs}(680)$  for all match-up points (total 29) points in Fig. 4A) are 19.80% and 31.24%. However, Dall'Olmo and Gitelson (2005) proposed that some noise caused by the atmosphere can be eliminated by bands ratio. This suggestion is consistent with our results, described below. The maximum and minimum relative error between measured and atmospheric correction results for  $R_{rs}(745) / R_{rs}(680)$  are significantly decreased, namely -5.18% for 49.75 µg/L C<sub>chl-a</sub> and -23.83% for 9.77 µg/L C<sub>chl-a</sub>. The RMSP of the measured and atmospheric correction result  $[R_{rs}(745) / R_{rs}(680)]$  for all match-up points is decreased to 17.61%.

#### 3.2.2. Validation of NIR-red two-band algorithm

The validation of the  $C_{chl-a}$  NIR-red two-band algorithm for matchup points shows that the performance of the  $C_{chl-a}$  NIR-red two-band algorithm is encouraging (Fig. 4A). The RMSPs between measured and retrieved  $C_{chl-a}$  are 47.12%, 36.60% and 46.55% on 08/01, 05/13 and 05/12, respectively. The minimum relative errors between measured and retrieved  $C_{chl-a}$  are -0.28%, 1.55% and 3.01% on 08/01, 05/13 and 05/12, respectively. The maximum relative errors between measured and retrieved  $C_{chl-a}$  are 43.90%, 62.11% and 42.80% on 08/01, 05/13 and 05/ 12, respectively. The distribution of  $C_{chl-a}$  in Taihu Lake, mapped by synchronous GOCI image data, also suggests that the regional empirical  $C_{chl-a}$  NIR-red two-band algorithm is promising as a means to retrieve  $C_{chl-a}$  for GOCI data (Fig. 4B). High  $C_{chl-a}$  was observed in Meiliang Bay, Zhushan Bay and Gonghu Bay. High  $C_{chl-a}$  was also observed in



Fig. 6. Hourly variation of C<sub>chl-a</sub> during daytime hours from August 6 to August 9 for different lake segments.

Southwestern Lake. These are almost consistent with the traditional in situ observations (Huang, Li, Yang, Sun, et al., 2014).

# 4. Results

# 4.1. Spatial variations of C<sub>chl-a</sub> in Taihu Lake mapped by GOCI data

Using the regional  $C_{chl-a}$  NIR-red two-band algorithm, we obtained hourly scale maps of C<sub>chl-a</sub> in Taihu Lake for four consecutive days (Fig. 5). There are large areas of submerged and emergent aquatic plants in Eastern Lake and particularly in Eastern Bay (Ma, Duan, Gu, & Zhang, 2008). The retrieved C<sub>chl-a</sub> in Eastern Lake and Eastern Bay should not be viewed as accurate due to the influence of seasonal water plants and bottom reflectance (Hu et al., 2010; Ma et al., 2008). Thus, according to the processing method of Hu et al. (2010), Eastern Lake and Eastern Bay were excluded in our analysis so as not to affect our analysis for other segments (Fig. 1A). A diurnal-spatial variation of C<sub>chl-a</sub> in Taihu Lake is clear. It can even be said that there is a clear hourly-spatial variation of C<sub>chl-a</sub> in Taihu. The hourly scale maps of C<sub>chl-a</sub> in the lake can show the dynamics of phytoplankton, including the vertical and horizontal directions. The areas in which  $C_{chl-a}$  was greater than 200  $\mu$ g/L were considered as phytoplankton bloom areas according to the longterm in situ measured  $C_{chl-a}$  and investigation. This is feasible and convenient to discuss the distribution and dynamics of phytoplankton bloom in our study periods.

Phytoplankton bloom was observed in Southwestern Lake on August 6 at 8:28 in the morning (Fig. 5A). The phytoplankton bloom became increasingly serious from 8:28 to 9:28 a.m. In the next hour (9:28 to 10:28), however, the surface accumulated phytoplankton bloom wore off significantly and disappeared in the next two hours (10:28 to 12:28). The  $C_{chl-a}$  in Southwestern Lake remained lower from 13:28 to 15:28. The  $C_{chl-a}$  in Central Lake went through a process from low to high and back to low. The  $C_{chl-a}$  in Northwestern Lake (including Zhushan Bay and Meiliang Bay) maintained a high level, and surface accumulated phytoplankton bloom appeared in some areas in Northwestern Lake.

After 17 h, or the next morning, a large area of surface accumulated phytoplankton bloom appeared in Northwestern Lake. The surface accumulated phytoplankton bloom covered almost the entire Zhushan Bay and Meiliang Bay at that point (Fig. 5B, 8:28). The duration of this floating phytoplankton bloom was very short. A large area of surface accumulated phytoplankton bloom in the Northwestern Lake disappeared, except in Zhushan Bay and in the waterfront of Western Lake, during the next 5 h (9:28 to 15:28). However, high  $C_{chl-a}$ , in Central Lake and northwestern area of Xishan Island, eventually formed a floating phytoplankton bloom in Central Lake (Fig. 5B, 15:28). From 9:28 to 13:28 on August 7, a patchy surface accumulated phytoplankton bloom appeared in Southwestern Lake (Fig. 5B, 11:28; 12:28). This

phytoplankton bloom was hard to monitor from the high time resolution satellite data.

On August 8, in contrast to the large area of floating phytoplankton bloom on the morning of August 7, a small surface accumulated phytoplankton bloom appeared in Zhushan Bay and in the waterfront of Western Lake (Fig. 5C, 8:28). At noon (Fig. 5C, 12:28), a high  $C_{chl-a}$  presented in Center and Southwestern Lake. However, this high  $C_{chl-a}$  did not form a floating phytoplankton bloom in Southwestern Lake, but a small floating phytoplankton bloom formed in Central Lake (Fig. 5C, 13:28 to 15:28).

The small floating phytoplankton bloom in Zhushan Bay, in the waterfront of Western Lake and in Central Lake (Fig. 5C, 15:28) aggregated into a large area of surface accumulated phytoplankton bloom in the morning of August 8 (Fig. 5D, 8:28). This floating bloom area became even bigger in the next hour. However, the floating bloom area then shrank, although there was still a big area of floating phytoplankton bloom in Central Lake from 12:28 to 13:28.

# 4.2. Hourly variation of C<sub>chl-a</sub> for each lake segment

Fig. 6 shows the temporal distributions of pixel average  $C_{chl-a}$  for each Taihu Lake segment from GOCI observations. There is an apparent difference during the period of August 6 to August 9. The significant increase of  $C_{chl-a}$  in Northwest Lake and Meiliang Bay in the morning from August 7 to August 9 corresponds to a large area of phytoplankton bloom in Fig. 5 (Fig. 5B, D, 8:28). The  $C_{chl-a}$  in Northwest Lake and Meiliang Bay maintains a high concentration from August 6 to August 9. Thus, these two segments have a high capacity to form floating phytoplankton bloom in subsequent days. The level of  $C_{chl-a}$  in Central Lake is relatively low, except on August 9, which is caused by the appearance of floating phytoplankton bloom. Similarly, high  $C_{chl-a}$  in Southwest Lake is accompanied by the appearance of floating phytoplankton bloom on the morning of August 6.



**Fig. 7.** Comparison of GOCI-derived  $C_{chl-a}$  and measured  $C_{chl-a}$  for match-up points from August 6 to August 9.



**Fig. 8.** Comparison of GOCI-derived  $C_{chl-a}$  and in situ buoy-measured  $C_{chl-a}$  during the daytime, August 6 to August 9. The points in the circle deviated from the buoy monitor.

# 5. Discussion

#### 5.1. Evaluation of capture capability of GOCI for C<sub>chl-a</sub>

In order to test the capture capability of  $C_{chl-a}$  in Taihu Lake by GOCI satellite image, a comparison of GOCI- retrieved  $C_{chl-a}$  with in situ match-up points and buoy measurement of  $C_{chl-a}$  was conducted. Fig. 7 shows a comparison of retrieved  $C_{chl-a}$  and in situ measurement of match-up points from August 5 to August 9. The RMSPs between measured and retrieved  $C_{chl-a}$  are 57.00%, 49.70%, 8.9%, 5.7% and 3.7% on 08/05, 08/06, 08/07, 08/08 and 08/09, respectively. The  $C_{chl-a}$  retrieved by GOCI was consistent with the in situ measured  $C_{chl-a}$ , which means that GOCI could reliably map the  $C_{chl-a}$  in Taihu Lake from 08/05 to 08/09.

How well were the dynamics of phytoplankton captured by GOCI? The diurnal variation of the GOCI-derived C<sub>chl-a</sub> was a good match with the buoy-observed data, except on August 8 (Fig. 8, marked by ellipse), when the retrieved  $C_{chl-a}$  was much lower than that observed by buoy. However, the retrieved  $C_{chl-a}$  was much higher than observed by buoy when phytoplankton bloom occurred on August 9 (8:28 and 9:28). There are several potential reasons for this anomaly. The observation depths of the buoy were 0.4 and 0.8 m (middle of water column), but the observation depths of the satellite were different. The difference between satellite and buoy observations will thus become evident when the vertical stratification of  $C_{chl-a}$  is very obvious. The diffuse attenuation coefficient in Taihu Lake is very large (Huang et al., 2009; Zhang, Qin, & Chen, 2005), e.g., the diffuse attenuation coefficient (at 490 nm) in summer can reach 6.8  $\pm$  3.7 m<sup>-1</sup> (mean value  $\pm$  standard deviation). Thus, the water constitutes in the deep water level are hard to be monitored in Taihu Lake. Consequently, the underestimation of  $C_{chl-a}$  on August 8 may result from the accumulation of phytoplankton in the middle of the water column, which GOCI may not fully detect. Fig. 5C (10:28 to 13:28) also shows that the high  $C_{chl-a}$  that appeared at 13:28 was the buoyant phytoplankton from the middle of the water column (there are no floating trails on the surface water). The overestimation of  $C_{chl-a}$  on August 9 (8:28 and 9:28) may occur because the phytoplankton is floating on the surface, and there is relatively little phytoplankton in the middle of the water column. Another possible reason for this overestimation is that the buoy (YSI) observed  $C_{chl-a}$  is only for algae (chlorophyll-a in the cyanobacteria was not fully detected by buoy) (Seppälä, Ylostalo, & Kaitala, 2007; Simis, Huot, Babin, Seppälä, & Metsamaa, 2012), and the GOCI-derived  $C_{chl-a}$  is for all phytoplankton.

# 5.2. Factors influencing distribution of C<sub>chl-a</sub>

#### 5.2.1. Growth of phytoplankton affected by nutrients and temperature

It is believed that phytoplankton blooms in Taihu Lake are primarily driven by nutrient levels but are also modulated by meteorological conditions (Hu et al., 2010; Huang, Li, Yang, Sun, et al., 2014; Wilhelm et al., 2011). The in situ measurement data of nutrients show that the total nitrogen (TN) and phosphorus (TP) are  $1.866 \pm 0.673$  mg/L (min-max: 1.013-5.135 mg/L) and  $0.156 \pm 0.076$  mg/L (min-max: 0.029-0.306 mg/L) from August 5 to August 9. The mean ratio of total nitrogen to total phosphorus (TN:TP) is  $15.171 \pm 8.341$  (min-max: 7.569-34.874 mg/L). The  $C_{chl-a}$  has a significant positive correlation to total nitrogen and total phosphorus, and the Pearson correlation coefficients are 0.79 and 0.62 (p < 0.001) (Fig. 9A). This means that the influence of total nitrogen on  $C_{chl-a}$  may stronger than that of total phosphorus on  $C_{chl-a}$  during our in situ measurement. This result is similar to that of Paerl et al. (2011). The result, which shows that  $C_{chl-a}$  is negatively correlated to TN:TP (Pearson correlation coefficients are -0.40, p < 0.005) (Fig. 9B), corresponds to the results of laboratory tests (Kim, Hwang, Shin, An, & Yoon, 2007). These indicate that both total nitrogen and total phosphorus should be reduced to controlled phytoplankton bloom in Taihu Lake. The negative correlation between  $C_{chl-a}$  and water temperature (Pearson correlation coefficients are -0.56, p < 0.001) (Fig. 9B) indicates that, relative to high temperature, low temperature in summer is much more suited to the phytoplankton growth in Taihu Lake. This may explain the large number of phytoplankton blooms that occurred in the morning with relatively low temperature (Fig. 5).

#### 5.2.2. Drift of phytoplankton bloom affected by hydrodynamic force

In order to characterize the hydrodynamic effect on the phytoplankton distribution, we can reasonably assume that the limitation of TN and TP to phytoplankton growth is minimal at the hourly scale, and that the hourly variation of phytoplankton is the result of modulation by meteorological conditions. The long-distance movement of a large area of phytoplankton on the water surface is rare within a short time period (Fig. 5). This phenomenon was also reported by Reynolds (2006) from dynamic simulation. It may result from low flow velocity. The horizontal laminar flow velocity is not in a fixed direction and is almost less than 10 cm/s from August 6 to August 9 (Fig. 10). This (Fig. 10A, B) indicates



Fig. 9. A is the relationship between  $C_{chl-a}$  and nutrients. B is the relationship between  $C_{chl-a}$  and temperature, TN:TP.



Fig. 10. Buoy-measured flow velocities for different directions and floors (A, B), East > 0 means the direction of flow is east, North > 0 is the same mean with East > 0; and West < 0 means the direction of flow is west, South < 0 is the same mean as West < 0. C is also the wind direction measured by buoy.

that, under this hydrodynamic condition, the phytoplankton can move at most 8640 m for four days (calculated by the maximum flow velocity = 10 cm/s and four days). Thus, the phytoplankton bloom in Zhushan Bay and Meiliang Bay on August 7 had not moved from the phytoplankton in Southwestern Lake on August 6 though laminar flow in the middle of the water column. Fig. 10C shows the wind direction. The flow velocity and direction of water surface are determined primarily by wind speed (within 10 m of the water surface). The floating phytoplankton bloom in Zhushan Bay and Meiliang Bay also did not come from the phytoplankton in Southwestern Lake though the surface current, although the surface current may be much bigger than laminar flow in the middle of the water (Fig. 5, showing that there is almost no floating trail of phytoplankton on the surface water). Wind-driven horizontal drift can contribute to the distribution of phytoplankton bloom, but was not the primary reason for the short-term phytoplankton bloom.

# 5.2.3. Floating and sinking of phytoplankton affected buoyancy and fluctuation

*Cyanophyta* were the dominant phytoplankton species in Taihu Lake in the course of this in situ measurement (*cyanophyta*: 97.04%  $\pm$  3.24%; *chlorophyta*: 0.79%  $\pm$  1.29% and *bacillariophyta*: 2.17%  $\pm$  2.47%). Unlike other phytoplankton species, *M. aeruginosa* (one kind of blue algae) can rapidly alter their buoyancy to change their vertical distribution (Sellner et al., 2003). Previous studies have shown that the vertical flow velocity can also significantly affect the vertical distribution of *M. aeruginosa* (Wallace et al., 2000; Fennel & Boss, 2003; Carstensen et al., 2007). Our GOCI observation demonstrated that wind-induced hydrodynamic mixing can carry the *M. aeruginosa* downward. As a result, surface bloom disappeared gradually on August 7 (8:28 to 10:28) and August 9 (9:28 to 15:28) (Fig. 5B, D). The wind threshold value for the disappearance of floating blooms differs among lakes, however, due to differences in depth and morphology, and to the parameters of the lake and the buoyancy of phytoplankton (Binding et al., 2011; Hunter et al., 2008a, 2008b; Webster & Hutchinson, 1994; Wynne et al., 2010). Field (Cao, Kong, Luo, & Shi, 2006) and satellite (MODIS data) (Huang, Li, Yang, Sun, et al., 2014) observations suggest that the phytoplankton bloom disappears when diurnal wind speed exceeds approximately 4 m/s. Huang et al. (2014b) has also described an empirical relationship between wind threshold value  $(W_t)$  and dynamic ratio  $(R_{dr}, is the ratio$ between square root of lake area and depth),  $W_t = 4.8679(R_{dr})^{-0.1}$ , according to observations at different lakes. However, in fact, wind threshold value is not a constant in Taihu Lake according to the observation wind speed (Fig. 11A) and GOCI-derived results from August 7 (8:28 to 10:28) and August 9 (9:28 to 15:28). For instance, the floating phytoplankton bloom began disappearing after 8:28 on August 7 with 5.1 m/s wind speed. However, the wind speed was 4.2 m/s after 9:28 on August 9. Additionally, no large area of floating phytoplankton bloom occurred on the morning of August 8 for high wind speed (7.8 m/s) at 21:15 on August 7. The different wind threshold values correspond to different duration times. Thus, the duration time of wind is another important factor. Fig. 11B shows the vertical flow velocity measured by buoy between -1.7 cm/s and 2.1 cm/s. This velocity is bigger than the buoyancy of M. aeruginosa (0.9 cm/s) calculated by Huang et al. (2014a), and easily pushes the algae (with 0.9 cm/s buoyancy) down.

# 6. Conclusions

Based on the GOCI-derived  $C_{chl-a}$ , in situ measured water quality parameters and buoy-measured  $C_{chl-a}$ , some major findings can be made from this study. First, the NIR-red two-band algorithm and 6S model with real-time atmospheric humidity and pressure can be applied to retrieve  $C_{chl-a}$  in Taihu Lake by GOCI data. Second, GOCI-derived  $C_{chl-a}$  can capture the spatial variation and dynamic characteristics of  $C_{chl-a}$  and phytoplankton bloom in the lake. Third, nutrients are required for phytoplankton bloom in the lake; the vertical movement induced by both phytoplankton buoyancy and fluctuation govern the hourly spatial



Fig. 11. A is the wind speed measured by buoy from August 6 to August 9. B is buoy-measured flow velocities for different floors, Up > 0 means the direction of flow is an upwelling current, and Down < 0 means the direction of flow is a downwelling current.

and temporal variation of phytoplankton, and the horizontal movement of phytoplankton is significant over the long term, but spatially and temporally limited in the short term.

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